

Running title: Trophic niches among tropical sharks

Interspecific variation in trophic niche metrics among euryhaline and coastal elasmobranchs in northern Australia

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1 **Abstract**

2 Tropical elasmobranchs are thought to play significant roles in many coastal and river
3 ecosystems, yet few studies have explored their trophic ecology. We investigated the
4 trophic niches of seven elasmobranch species in northern Australia, using stable
5 carbon ($\delta^{13}\text{C}$) and nitrogen ($\delta^{15}\text{N}$) isotopes (SI) and fatty acid (FA) markers taken
6 from muscle tissues sampled during the tropical monsoonal wet and dry seasons. Both
7 SI and FA metrics suggested significant niche partitioning between the species; with
8 two distinct guilds apparent from the SIs: a marine foodweb based on macroalgae and
9 seagrass, and an estuarine/freshwater foodweb with a seston base. Fatty acid profiles
10 showed contributions from both macroalgal and seston based food webs. Varying
11 degrees of interspecific niche distinction were apparent: *Glyphis garricki*, *G. glyphis*
12 and *Himantura dalyensis* had FAs from vascular plants and diatoms/macroalgae
13 which separated them from *Carcharhinus leucas*, *C. amboinensis* and *Rhizoprionodon*
14 *taylori* that had FA markers associated with marine dinoflagellates. Niche metrics
15 differed between methods, where for example, *G. garricki* had the smallest SI niche
16 area yet the largest FA niche space, but a generally low probability of overlap with
17 other species. Differences in the niche metrics between SI and FA are likely due to
18 their disparate turnover times. Our results suggest that coastal and euryhaline
19 elasmobranchs provide trophic connections between tropical marine, estuarine and
20 freshwater ecosystems.

21

22 **Key words:** niche metrics, sharks, *Glyphis*, *Carcharhinus*, *Rhizoprionodon*,
23 biotracers, stable isotopes, fatty acids

24

25

26 **Introduction**

27 Resource partitioning, whereby species segregate habitats and dietary items over
28 space and time (Ross 1986), is fundamental to our understanding of coastal ecosystem
29 structure and function (Kitchell et al. 2002). Elasmobranchs (sharks and rays) are
30 demographically vulnerable predators that are thought to play key roles in aquatic
31 food webs (Stevens et al. 2000, Dulvy et al. 2014). As middle-order or apex predators,
32 elasmobranchs can provide ecosystem stability by influencing the abundance and
33 health of prey species at multiple trophic levels, and connect otherwise distinct food
34 webs (Rooney et al. 2006, Heithaus et al. 2013).

35

36 Niche theory suggests that species have a set of specific resource requirements that
37 form their unique space, an “ n – dimensional niche hypervolume” (Hutchinson 1957),
38 which may be measured and compared among locations, seasons and species. In some
39 cases, species can share resources to produce niche overlap. Whether such overlap
40 leads to competition is dependent on the spatial and temporal extent of shared
41 resource use. In the case of trophic overlap, sympatric species may target different
42 prey items (Yick et al. 2011), and/or have temporal or spatial differences in
43 consumption that negate competitive interactions (Ross 1986). Sympatric
44 elasmobranchs have been found to have variable trophic overlap, ranging from high
45 (e.g. sympatric batoids, Banded Stingaree *Urolophus cruciatus* and Tasmanian
46 Numbfish *Narcine tasmaniensis* (Yick et al. 2011)) to low (e.g. Caribbean Reef Shark
47 *Carcharhinus perezi*, Nurse Shark *Ginglymostoma cirratum* and Southern Stingray
48 *Dasyatis americana* (Tilley et al. 2013)) in a range of elasmobranch assemblages
49 (Papastamatiou et al. 2006, Heithaus et al. 2013).

50

51 Euryhaline elasmobranchs are capable of tolerating a range of salinities and may
52 complete different life history stages in marine, estuarine or freshwater habitats with
53 tropical regions having the most diversity of euryhaline elasmobranch species
54 (Lucifora et al. 2015). Due to their reliance on rivers and estuaries, euryhaline
55 elasmobranchs must adapt to fluctuations in salinity and turbidity over tidal and
56 seasonal cycles, both in terms of their own physiology, and the abundance of potential
57 prey items. Such environmental changes are particularly extreme in tropical rivers due
58 to the variation between high-flow monsoonal ‘wet’ seasons and long low-rainfall
59 ‘dry’ seasons (Douglas et al. 2005, Warfe et al. 2011). Trophic resource use and
60 partitioning within tropical elasmobranch assemblages is poorly understood at
61 present, although recent studies have indicated the existence of overlap among
62 juvenile *Carcharhinus amboinensis* and *C. leucas* in northern Australia (Tillett et al.
63 2014), and among *C. leucas* and *Pristis pristis* in Western Australia (Thorburn et al.
64 2014).

65
66 Analysis of biochemical tracers in animal tissue, such as stable isotopes (SI) and fatty
67 acids (FAs), is an effective approach for estimating time-integrated trophic niches of
68 aquatic species (Hussey et al. 2011, Jackson et al. 2011, Pethybridge et al. 2014). The
69 predictable fractionation of stable isotopes (e.g., $\delta^{13}\text{C}$, $\delta^{14}\text{N}$) through food webs,
70 allows isotopes to be used for a variety of applications, including estimation of niche
71 area, resource partitioning and dietary overlap. Fatty acids can also be traced through
72 food webs, and are particularly useful for detecting trophic markers via essential FAs
73 (EFAs) that can only be obtained through dietary sources (Iverson 2009). These EFAs
74 are synthesized by different primary producers such as dinoflagellates, diatoms or
75 algae, after which they are accumulated up the food chain (Dalsgaard et al. 2003,

76 Parrish et al. 2013). Fatty acids often have faster turnover rates than isotopes (e.g. 14
77 weeks in shark muscle tissue (Beckmann et al. 2014), cf. 6 – 12 months for $\delta^{13}\text{C}$ and
78 $\delta^{15}\text{N}$, (Malpica-cruz et al. 2012, Hussey et al. 2012), and have been used to reveal
79 physiological and environmental changes over relatively fine scales, such as seasonal
80 differences in the EFA trophic biomarker omega 3 [$\omega 3$] / omega 6 [$\omega 6$] (Jayasinghe et
81 al. 2003, Pethybridge et al. 2015).

82

83 Northern Australia provides an excellent setting for studying tropical elasmobranch
84 trophic ecology due to the relatively pristine state of the tropical estuarine and riverine
85 ecosystems (Warfe et al. 2011) and a high diversity of sympatric elasmobranch
86 species (Last & Stevens 2009). This includes euryhaline (*C. leucas*, *Glyphis garricki*,
87 *G. glyphis*, *Himantura dalyensis*, *P. pristis*) and coastal (*C. amboinensis* and
88 *Rhizoprionodon taylori*) species. In the current study, we use SI and FA biochemical
89 tracers to evaluate the extent of trophic resource overlap or partitioning among seven
90 euryhaline and coastal elasmobranchs during the wet and dry season in northern
91 Australia. We also use these tracers to identify the basal sources of food webs
92 (marine, estuarine, fresh water) and apply biochemical niche metrics (Jackson et al.
93 2011, Swanson et al. 2015) to determine the extent of dietary overlap within and
94 among species across seasons.

95

96 **Methods**

97 *Sample collection and preparation*

98 Seven species of euryhaline and coastal elasmobranchs were collected in the South
99 Alligator River, Northern Territory (NT), Australia from March 2013 to July 2014
100 (Table 1, Fig. 1). This time period included the two broad seasons, the ‘wet’ and

101 'dry'. The wet is characterized by high humidity, rainfall and temperature from
102 November to March followed by decreases in these parameters during the dry season
103 from April to October. Linked to the variation in rainfall, salinity also varies
104 throughout the seasons. Salinity was measured with an YSI (Xylem, USA) at each site
105 (Fig. 1), during each sampling event. The mouth and coastal region of the river ranged
106 from 32.4 – 34.5‰ (dry) and 17.1–24.7‰ (wet) whereas the mid-lower region of the
107 river ranged from 18.3 – 28.1‰ (dry) and 0.4 – 11.3‰ (wet).

108

109 *Rhizoprionodon taylori* and *Carcharhinus amboinensis* were captured with baited line
110 in the lower estuary and coastal region of the South Alligator River, Kakadu National
111 Park (Fig. 1). *Glyphis garricki*, *G. glyphis*, *C. leucas*, *Pristis pristis* and *Himantura*
112 *dalyensis* were caught in mid-lower estuarine reaches with a combination of fishing
113 methods including 4 – 6 inch gill nets and baited lines. A 5 mm biopsy punch (Stiefel,
114 USA) was used to collect muscle tissue from the caudal peduncle within 5 minutes of
115 capture, with the tissue sample immediately placed on liquid nitrogen (-196°C) for
116 preservation and initial storage in the field. Within one week, tissues (mean wet
117 weight 0.02 ± 0.01 g) were transferred to a -20°C freezer, and then later freeze-dried
118 (Alpha 1-4 LSC, Christ, Germany) at -20°C for 21 hours and then -30°C for three
119 hours. To avoid degradation of the sample from defrosting and refreezing, frozen
120 muscle samples were dissected in the freezer to remove dermal layers and as much
121 connective tissue as possible. Where sample tissue masses were low, tissue was only
122 used for SIA and not FA analysis (Table 1).

123

124 *Stable Isotopes*

125 Muscle tissue was rinsed in milli-Q water and sonicated to remove excess urea as per
126 Kim & Koch (2011). Tissues were then freeze-dried to a constant weight and
127 pulverized using a combination of micro-scissors and micro-pestle or a coarse pestle
128 and ceramic mortar. A subset of material was weighed to 400 – 1000 ug in tin cups
129 for combustion in a Sercon Europa EA-GSL elemental analyzer (Sercon Ltd, UK),
130 which were then gas analyzed with a Sercon Hydra 20 – 22 – isotope ratio mass
131 spectrometer (Sercon Ltd, UK) at the Australian Rivers Institute, Griffith University.
132 The international standards used to determine the relative $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ were Peedee
133 Belemnite Carbonate and Atmospheric Nitrogen. We did not use mathematical
134 models to correct $\delta^{13}\text{C}$ values for lipids as lipid content in all samples were deemed to
135 be low as inferred by ratios of total organic carbon to total organic nitrogen (C:N)
136 being < 3.5% (Post et al. 2007) and by the low lipids inferred from total FAs (Every et
137 al. In press).

138

139 Trophic positions (TP) for each species were calculated using the narrowing
140 discrimination with increasing dietary $\delta^{15}\text{N}$ values approach (Hussey et al. 2014).
141 This combined two models: a meta-analytical Bayesian model of diet discrimination
142 factors, and the $\delta^{15}\text{N}$ values with a von Bertalanffy growth model. This works on the
143 concept that as trophic position (TP) increases, the amount of dietary discrimination
144 factors (the increase of $\delta^{15}\text{N}$ at each TP) is not constantly added, but decreased at each
145 TP (Hussey et al. 2014). To calculate this, an inverse relationship of dietary
146 discrimination factors to $\delta^{15}\text{N}$ was determined, based on known discrimination factors
147 in Hussey et al. (2014), which was then applied to data used in this study. Popeye
148 mullet (*Rhinomugil nasutus*, $\delta^{15}\text{N}$ 6.62 ± 1.19 , trophic position 2.92 (Froese & Pauly
149 2015)) was used as the baseline consumer collected at the same location and over the

150 same time period as the elasmobranch tissue samples. The following equation was
151 used with values from Hussey et al. (2014) meta analysis:

152

153
$$TP = \frac{\log(\delta^{15}N_{lim} - \delta^{15}N_{base}) - \log(\delta^{15}N_{lim} - \delta^{15}N_{TP})}{k} + TP_{base}$$

154

155 where TP_{base} is the trophic position of baseline consumer (*R. nasutus*), $\delta^{15}N_{TP}$ is the
156 consumer $\delta^{15}N$ value and k is calculated from $\beta_0 = 5.92$ [4.55, 7.33] and $\beta_1 = -0.27$ [-
157 0.41, -0.14] after (Hussey et al. 2014):

158
$$k = -\log \frac{(\beta_0 - \delta^{15}N_{lim})}{\delta^{15}N_{lim}}$$

159 where term:

160
$$\delta^{15}N_{lim} = \frac{-\beta_0}{\beta_1}$$

161

162 *Fatty acids*

163 Lipid was extracted using the modified Bligh and Dyer (1959) method, which utilizes
164 an overnight one-phase extraction process. The extracted lower layer was
165 concentrated, blown down with nitrogen gas and dried to a constant weight.
166 Approximately half of the total lipid extract was transmethylated (Parrish et al. 2015)
167 to liberate the FAs from the lipid backbone. After the solution was prepared, 0.2 ml
168 was injected into an Agilent Technologies 7890B gas chromatograph (GC) (Palo Alto,
169 California, USA) equipped with an Equity-1 fused silica capillary column (15 x 60.1
170 mm i.d. and 0.1 mm film thickness), a flame ionization detector, a splitless injector
171 and an auto-sampler. Peaks were quantified using Agilent Technologies ChemStation
172 software (Palo Alto, California, USA). Confirmation of peak identifications was by

173 GC-mass spectrometry (GC-MS), using a column of similar polarity to that described
174 above and a Finnigan Thermoquest DSQ GC-MS system.

175

176 *Niche area calculations*

177 A SI Bayesian ellipse model (SIBER) using the R package Stable Isotopes Analysis in
178 R (SIAR) was used to measure two different metrics of isotopic niche areas, convex
179 hull total area (TA) (Layman et al. 2007, Jackson et al. 2011) and the stable isotope
180 ellipse area corrected for sample size (SEA_c) (Jackson et al. 2011). Although TA in a
181 $\delta^{15}\text{N} - \delta^{13}\text{C}$ biplot is strongly affected by sample size (Jackson et al. 2011, Syväranta
182 et al. 2013) it was calculated to explore species variation and the potential overlap of
183 individuals (Layman et al. 2012, Heithaus et al. 2013). Drawing standard ellipse areas
184 (SEA) in isotopic space ($\delta^{13}\text{C}$ and $\delta^{15}\text{N}$) that incorporated 95% of the sampled
185 individuals, is considered to provide a more accurate measure of niche width than TA
186 and is less biased by sample size (Jackson et al. 2011, Syväranta et al. 2013). To
187 further deal with small, uneven sample sizes, we applied a sample size-correction
188 measure, calculated from the covariance matrix determined via Bayesian inference to
189 enable the area to be measured (SEA_c). Confidence intervals of the ellipse size are
190 calculated from Bayesian likelihoods, where by the probability of one ellipse being
191 bigger than the other can be determined by the uncertainty in the probable values.
192 Overlaps of (SEA_c) were calculated by establishing the percent difference between
193 the ellipses. Only *C. leucas*, *G. garricki* and *R. taylori* had ellipses analyzed as *n*
194 values were over 30, which is considered optimum for these analysis (Syväranta et al.
195 2013).

196

197 Fatty acid niche space and overlap was calculated using a multivariate extension on
198 the bivariate approach of Jackson et al. (2011). Here, Bayesian priors are used to
199 determine the hyper-volume and the probability of finding one species in another's
200 niche space (Swanson et al. 2015). To increase the confidence of this model the
201 amount of FAs used was reduced as the numbers of individuals were considered to be
202 too low for the number of variables (M. Lysy, personal communication, 6/11/2015).
203 Therefore for ~30 individuals of each species, five FAs were used and only *G.*
204 *garricki*, *C. leucas* and *R. taylori* were analyzed. Essential FAs were selected since
205 they are largely accumulated through diet. The most abundant EFAs were also used.
206 Each FA was then represented as an axis, calculated in FA hyperspace and projected
207 onto a two dimensional plot as a probabilistic projection to display FA niche space.
208 Overlap was calculated based on the probability that one species would be found in
209 the niche of another species and a mean value of niche space calculated.

210

211 *Statistical Analysis*

212 Analysis of variance (ANOVA) was used to determine differences in stable isotope
213 tracers among species and seasons. For *C. leucas* an Analysis of Covariance
214 (ANCOVA) was run with total length (TL) as a covariate to account for the effects of
215 length on isotopic values. Assumptions of normality, variance, homogeneity of slopes
216 and collinearity were tested through visual inspection of boxplots, calculating
217 Leverage, Residuals and Cook's D and graphed in a Residual vs Fitted, Normal Q-Q
218 and Residual vs Fitted values plots prior to accepting the models. No transformation
219 was necessary.

220

221 Fatty acids were converted to a percentage and those with group means less than 0.5%
222 were not included in statistical analyses. To evaluate interspecific differences in FA
223 profiles a one-way semi-parametric permutation multivariate analysis of variance
224 (PERMANOVA) routine was applied to a Euclidean distance matrix, using a
225 minimum of 9900 unique permutations. To confirm which FAs were most responsible
226 for observed interspecific differences an analysis of similarity (SIMPER) was
227 conducted. To test the effects of season on FA profiles, a second PERMANOVA was
228 run only on *G. garricki*, as seasonal-based n-values were low in other species (Table
229 1). To compare FA profiles between species and season, a Principal Coordinate
230 Analysis (PCO) was constructed from non-data arranged into a difference matrix
231 based on Euclidean distances. Analysis of covariance was used to explore
232 interspecific differences in $\omega3/\omega6$ between *G. garricki*, *G. glyphis*, *C. leucas* and *R.*
233 *taylori*. As significant effects were found, we then used an ANCOVA with TL as a
234 covariate, between each shark's $\omega3/\omega6$ ratio to consider possible allometric effects.
235 This ratio has been linked to changes in environmental responses and may assist in
236 explaining interspecific differences.

237

238 Univariate analyses were performed in R using the R core packages (R Development
239 Core Team 2014), SIBER (Jackson et al. 2011) and one FA multivariate analysis used
240 the package NicheROVER to determine overlap between sharks (Swanson et al.
241 2015) whilst the other multivariate tests were performed in PRIMER (v6) (Clarke &
242 Gorley 2006).

243

244 **Results**

245 *Stable isotopes*

246 Muscle $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ values in seven elasmobranch species revealed two distinct
247 trophic guilds (Fig. 2). The most enriched values were in *C. leucas*, *R. taylori*, *C.*
248 *amboinensis* and *P. pristis* (Table 2, Fig. 2). The elasmobranch guild with the lowest
249 range of $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ values included *G. glyphis*, *H. dalyensis* and *G. garricki*
250 (Table 2, Fig. 2). When comparing all species there was a significant relationship
251 between $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ ($p > 0.01$, $F_{df} = 33.2_{1, 117}$, $R^2 = 0.22$) and between the ratio of
252 C:N and $\delta^{13}\text{C}$ ($p > 0.01$, $F_{df} = 14.81_{1, 117}$, $R^2 = 0.11$). Total length was not significant in
253 any species or isotope, with the exception of $\delta^{13}\text{C}$ values in *C. leucas* ($p > 0.01$, $F_{df} =$
254 $10.29_{1, 30}$, $R^2 = 0.25$).

255

256 Based on $\delta^{15}\text{N}$ values, the calculated TP ranged from 3.2 ± 0.8 to 4.8 ± 0.9 and was
257 lowest in *G. glyphis* and highest in *C. leucas* (Table 2). Analysis of variance indicated
258 significant differences among species for $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$. The post hoc test for $\delta^{15}\text{N}$
259 showed no significant differences between *R. taylori* and *C. leucas*, and between *G.*
260 *garricki* and *G. glyphis*, whereas all others species differed significantly. For $\delta^{13}\text{C}$,
261 only *G. garricki* and *G. glyphis* were not significantly different from each other.

262 Although seasonal effects for $\delta^{13}\text{C}$ were found in *C. leucas*, the inclusion of TL as a
263 covariate negated the seasonal effect and revealed TL to be the only significant
264 variable (Table 3).

265

266 Of the three species that convex hull total area (TA) was calculated for, *C. leucas* was
267 the largest followed by *R. taylori* and *G. garricki*. *Carcharhinus leucas* overlapped
268 strongly with *R. taylori*, although some individuals were closer in niche space to *G.*
269 *garricki*. *Glyphis garricki* had the least overlap with other species: only the outliers of
270 *R. taylori* and *C. leucas* shared their TA. Convex hull total areas were larger in the dry

271 season than the wet season for the three sharks. *Rhizoprionodon taylori* and *C. leucas*
272 had the biggest differences between the wet and dry season.

273

274 Size-sample corrected standard ellipse areas (SEAC) of stable isotopes ranged from
275 6.86 to 18.46 (confidence intervals (CI) 97 – 99%) with *C. leucas* having the largest
276 area followed by *R. taylori* and *G. garricki* (Table 4, Fig. 1 in the Supplement 1).

277 *Carcharhinus leucas* and *R. taylori* SEAC overlapped and were clearly separated from
278 *G. garricki* (Table 4, Fig. 1 in the Supplement 1). There was no overlap between *G.*

279 *garricki*, *C. leucas* and *R. taylori*. However, *C. leucas* and *R. taylori* had a partially
280 overlapping SEAC. All species had seasonal differences in SEAC, with higher values
281 in the dry season for *G. garricki*, and lower dry season values in *R. taylori* and *C.*

282 *leucas* (Fig. 2, Table 4). Seasonal differences were also evident in the overlap of the

283 SEAC between *C. leucas* and *G. garricki* and, *G. garricki* and *R. taylori* with more

284 overlap in the dry than the wet season (Fig. 2, Table 4). The SEAC of *C. leucas* in the

285 dry season slightly overlapped with that of *G. garricki*, although there was little niche

286 overlap with *G. garricki* over the whole sampling period (Table 4, Fig. 3).

287

288 *Fatty acids*

289 Sixty-five FAs were identified in the elasmobranch muscle tissue, of which 31 were

290 detected with mean values above 0.5% for any one species (Table 1, 2, and Table 1 in
291 the Supplement 2). Relative abundance of saturated FAs (SFAs) were similar in *C.*

292 *leucas*, *G. garricki*, *G. glyphis* and *H. dalyensis* ranging from 27.5 ± 7.2 to $32.0 \pm 8.7\%$

293 but were higher in *R. taylori* and *C. amboinensis* (42.5 and $43.6 \pm 12.3\%$, respectively).

294 The monounsaturated FAs (MUFA) were relatively low in *G. glyphis*, *G. garricki* and

295 *R. taylori* (means ranging from 18.8% – 22.2%) compared to other elasmobranchs,

296 whose mean relative abundance ranged from 24.8% to 29.7%. Relative amounts of
297 polyunsaturated FAs (PUFA) were highly variable between species (means ranging
298 from 14.8% – 44.0%) with it lowest in *H. dalyensis* and highest in the *Glyphis*
299 species. Polyunsaturated FAs were similar between *C. leucas* and *R. taylori*
300 ($30.3 \pm 2.2\%$ and $30.9 \pm 3.4\%$, respectively) and were dominated by three EFAs:
301 20:4 ω 6, 22:6 ω 3 and 22:4 ω 6.

302

303 There were significant differences in FA profiles between species (PERMANOVA:
304 $F_{df} = 9.163$, $p < 0.01$) and there was no effect of season in the *G. garricki* FA profiles
305 ($F_{df} = 0.691$, $p = 0.60$). Elasmobranch FAs showed grouping between species, overlap
306 and some slight seasonal differences (Fig. 4). *Glyphis garricki* collected in the wet
307 season had similar FA profiles to those caught in the dry season. The greatest overlap
308 was between *G. glyphis* and *G. garricki* and one subgroup of *C. leucas*. The least
309 overlap was in *R. taylori*, separated by the SFAs: 17:0, 18:0 and 16:0. A large
310 subgroup of *C. leucas* was also separated from the other species by the FAs, 18:2c,
311 20:3 ω 9, 18:2b, 18:1 ω 9 and 20:2. Whereas the *Glyphis* spp. were largely separated by
312 EFAs: 20:3 ω 6, 22:6 ω 3, 22:4 ω 6, 22:5 ω 6 and 20:4 ω 6.

313

314 The highest mean ratio of ω 3/ ω 6 was in *R. taylori* and the lowest was in *G. glyphis*
315 ranging from (0.9 ± 1.4 – 0.4 ± 0.8) (Table 2). An ANOVA of these taxa revealed
316 significant interspecific differences ($p = 0.01$, $F_{df} = 7.23, 76$, $R^2 = 0.22$), however post
317 hoc tests found significant differences between *G. garricki* and *C. leucas*, and *G.*
318 *glyphis* and *C. leucas* and differences between *R. taylori* and the other sharks; *G.*
319 *garricki*, *G. glyphis* and *C. leucas*. Total length had a significant effect on ω 3/ ω 6

320 ratios in *C. leucas* ($p = 0.05$, $F_{df} = 4.21, 21$, $R^2 = 0.18$), *G. garricki* ($p = 0.02$, $F_{df} = 6.21,$
321 23 , $R^2 = 0.18$) but not in *R. taylori* or *G. glyphis*.

322

323 The largest FA niche size was *G. garricki* followed by *C. leucas* and *R. taylori* (Table
324 4, Fig. 5 and Fig. 3 in the Supplement 3). The chance of *G. garricki* being in another
325 species' niche space was low (0.6% for *R. taylori* and 13.2% for *C. leucas*) yet *C.*
326 *leucas* and *R. taylori* had larger probabilities for being in *G. garricki* niche space
327 (79.9% for *R. taylori* and 61.7% for *C. leucas*). Slight seasonal differences were
328 evident in *G. garricki* FA niche space with a 15% probability of difference between
329 the wet and dry (Table 4).

330

331 *Trophic niche metric comparisons*

332 Taxon and seasonal groupings were less prominent for FA profiles than for isotopic
333 profiles, which highlighted two distinct guilds. Niche metric calculations supported
334 these variations but there were differences between methods used, which meant they
335 were not directly comparable. Between the two SI niche metrics (SEA_C and TA), TA
336 was the largest and had the most overlap compared to SEA_C . Also the SEA_C overlap in
337 *G. garricki* was slightly larger during the wet than the dry, yet the reverse was true in
338 TA overlap. *Carcharhinus leucas* showed the greatest difference between SI and FA
339 niche metrics. In SEA_C , *C. leucas* had the most overlap with *R. taylori*, yet in FA niche
340 space *C. leucas* had the most overlap with *G. garricki* (Table 4). As FA niche space is
341 a probability projection rather than geometric (Swanson et al. 2015), different
342 probabilities can be obtained for the same two species as each calculation is the
343 likelihood of one species being in the others' niche space. For example *R. taylori* had a
344 high probability of being in *G. garricki* niche space but the reverse was not true. This

345 was quite different from SEAc where *G. garricki* and *R. taylori* had little chance of
346 sharing resources except in TA. Seasonal differences between *G. garricki* SI and FA
347 niche probability appeared similar with a 15% difference between wet and dry values
348 and the overlap difference in SI was 9.5%.

349

350 **Discussion**

351 Analysis of SI and FA niche metrics demonstrated significant differences in the
352 trophic resource use of elasmobranch species, with indications of resource
353 partitioning within the South Alligator River ecosystem and surrounding coastal
354 waters. Although SI showed clear separation between coastal/marine and euryhaline
355 elasmobranchs, the difference was less pronounced between FAs. This is likely due to
356 the faster turnover of FAs (Kirsch et al. 1998, Beckmann et al. 2014) compared to SIs
357 (Logan & Lutcavage 2010) and the larger number and biochemical roles of FAs.
358 Seasonal differences in biota are relatively common in tropical river systems due to
359 the large outflow of freshwater during the wet season changing the sources of energy
360 and nutrients, and increasing primary and secondary production (Winemiller & Jepsen
361 1998). However, our results showed seasonal change in all SI and FA niche metrics
362 but not in the individual biochemical tracers. We observed only slight seasonal
363 changes within *Glyphis garricki* FA niche space that again may be the result of the
364 faster turnover of FAs. However, seasonal differences in $\delta^{15}\text{N}$ and $\delta^{13}\text{C}$ were also
365 marginal, suggesting that *G. garricki* are consuming similar prey over both seasons.

366

367 Distinct differences in the isotope niche space of elasmobranchs appear to be largely
368 driven by $\delta^{13}\text{C}$, with several species (*Rhizoprionodon taylori*, *C. amboinensis*, *C.*
369 *leucas* and *Pristis pristis*) characterized by values that are typically reported in marine

370 environments (-14 – -16‰; Hussey et al. 2011, Munroe et al. 2015). $\delta^{13}\text{C}$ values for
371 *P. pristis* and *C. amboinensis* were similar to previous studies (Knip et al. 2011,
372 Thorburn et al. 2014), which may have been sourced from seagrass, whilst *C. leucas*
373 and *R. taylori* had values closer to marine macroalgae (Loneragan et al. 1997). In
374 contrast, *G. garricki*, *G. glyphis* and *Himantura dalyensis* had depleted $\delta^{13}\text{C}$ values
375 suggestive of estuarine/freshwater seston signatures (-18.8 – -23.2‰; Loneragan et al.
376 1997). Notably, large standard deviations in *G. glyphis* $\delta^{13}\text{C}$ values (compared to *G.*
377 *garricki*) may be indicative of a greater variation in dietary sources consumed by *G.*
378 *glyphis* and/or our relatively small sample size in this species.

379

380 In terms of trophic position (TP), the seven elasmobranch species were broadly
381 similar, based on the spread of $\delta^{15}\text{N}$ values; with *C. leucas* having the most enriched
382 and highest mean TP (4.8 ± 0.9). Although our analysis used juveniles, this value was
383 higher than adult TP (4.6 ± 0.2) and sub-adult TP (4.4 ± 0.3) *C. leucas* reported in
384 Mozambique (Daly et al. 2013) and Western Australia TP (~ 4.4) (Thorburn et al.
385 2014) and in the same range (~3.8 – 5.4) as those from South Africa (Hussey et al.
386 2014). This variation is likely due to the different methods of determining TP as well
387 as dietary and environmental variation (Peterson & Fry 1987). The method developed
388 by Hussey et al. (2014) provides more discrimination between predators and the
389 authors found that this method fits logically with TP calculated from prey type in
390 stomach content analysis. Moreover, the FA 18:1 ω 9 that is often linked to piscivory
391 and thus suggestive of a higher TP was also in relative high abundance in *C. leucas*
392 (Dalsgaard et al. 2003, Kelly & Scheibling 2012). Trophic position was not calculated
393 in the Queensland population of *R. taylori* (Munroe et al. 2015) however $\delta^{15}\text{N}$ results
394 were lower, which is possibly explained by different environmental conditions.

395 *Glyphis glyphis* appear to be consuming lower order, benthic prey as their low TP was
396 more similar to rays than sharks of this body size (Hussey et al. 2014) (~ maximum
397 size 260 cm (White et al. 2015)). *Glyphis garricki* had a higher TP than *G. glyphis*,
398 suggesting this species consumes higher order prey than its sympatric congener.

399

400 The ratio of ω_3/ω_6 can be used to help define a species' niche as higher ratios ($>$
401 ~ 1.5) signify a preference for marine resources whilst lower ratios ($< \sim 1.5$) indicate a
402 preference for freshwater resources (Martínez-Álvarez et al. 2005, Özogul et al.
403 2007). The elasmobranchs in the current study appeared to follow this pattern as *R.*
404 *taylori* (a marine species) had the highest ratio, whilst species collected upstream
405 tended to be lower. However, *R. taylori* was not significantly different from the other
406 species, which may have been due to their high intraspecific variation. Differences in
407 the ω_3/ω_6 ratio were found in teleost fish when they were acclimatizing to changing
408 saline conditions (Martínez-Álvarez et al. 2005), which may explain why *R. taylori*
409 has high variance given that they are consuming prey around the mouth of the river
410 and may experience lowered salinity. Ratios in *C. leucas* and *G. garricki* were
411 significantly different which may reflect a preference for differing salinities. This is
412 interesting, as significant differences between salinities have not previously been
413 recorded for elasmobranchs (Speers-Roesch et al. 2008), however further studies
414 would be required to confirm this in these species. The ω_3/ω_6 ratio may also be
415 related to growth and maturation in elasmobranchs, as a significant relationship was
416 found between TL and the ratio in *C. leucas* and *G. garricki*, which were mostly
417 juveniles. Uysal & Aksoylar (2005) found that ω_3/ω_6 ratios in teleost fish decreased
418 during sexual maturation.

419

420 Individual FAs indicate complexity in the trophic resource niches among and within
421 species. For example, the FA profiles of *Glyphis* spp. suggest links to both marine and
422 freshwater ecosystems due to the presence of 18:2 ω 6 and 20:5 ω 3 which are typically
423 associated with marine algae, vascular plants and mangroves (Parrish 2013). Although
424 *C. leucas*, shares niche space with the *Glyphis* spp., they may be utilizing more
425 freshwater sources because *C. leucas* also had FAs in common with *H. dalyensis* such
426 as 16:1 ω 7 which has been linked to mangroves and diatoms (Parrish 2013).
427 *Himantura dalyensis* also had a low abundance of 22:6 ω 3 (typically associated with
428 marine dinoflagellates (Parrish 2013)), suggesting that they utilize freshwater
429 resources. Additionally, the relatively high abundance of 20:4 ω 6 conforms with their
430 morphology and benthic feeding preference (Hall et al. 2006, Pethybridge et al. 2011).
431 Finally, *C. leucas* had more FAs in common with the *Glyphis* spp. than *R. taylori*
432 although *G. glyphis* had higher abundances of 20:4 ω 6 suggesting that there are links
433 between marine and estuarine/coastal prey (Hall et al. 2006, McMeans et al. 2013).
434 Thus, *C. leucas* FA profile suggests a greater intake of estuarine prey than marine,
435 contrary to SI values. *Rhizoprionodon taylori* had higher abundances of the marine
436 FA marker, 22:6 ω 3 compared to the other species, which would be expected given
437 their marine habitat preference. The two juvenile *C. amboinensis* FA profiles were
438 closest to *R. taylori*, and one of the dominant EFAs (20:4 ω 6) in *C. amboinensis* were
439 in abundances that may be found in estuaries.

440

441 Differences in SI values and FA profiles in *C. leucas* may be a result of maternal
442 influences, while differences in turnover between the two types of biotracers (SI, FA)
443 are likely to be most linked to temporal differences. The significant correlation
444 between TL and *C. leucas* $\delta^{13}\text{C}$ values, suggests that these sharks rely on different

445 food sources as they grow. Maternal signatures have been found in *C. leucas* in both
446 $\delta^{15}\text{N}$, $\delta^{13}\text{C}$ and FAs (Matich et al. 2010, Olin et al. 2011, Belicka et al. 2012),
447 however it is possible that size-related differences in $\delta^{15}\text{N}$ were not found in these
448 sharks because the size range was too small. Tissues in *C. leucas* have shown a
449 decline in isotopic values until they reach 110 – 130 cm TL (Matich et al. 2010)
450 whereas our maximum TL was 86 cm. The two *P. pristis* also had enriched SI values.
451 Considering they were juveniles and caught in the mid-section of the sampling sites
452 near some *G. garricki* who had lower $\delta^{13}\text{C}$, which may suggest a maternal influence.
453
454 As stomach contents analysis in *C. leucas* have found a broad range of species
455 including catfish (ariids), rays (batoids), and other carcharhinid sharks (Snelson et al.
456 1984, Tillett et al. 2014), a large trophic niche would be expected. This was true of SI,
457 however the size of their FA niche space was intermediate between *G. garricki* and *R.*
458 *taylori*. This pattern may indicate evidence of individual specialization as was found
459 in Florida, USA (Matich et al. 2011), diet switching (Matich & Heithaus 2014) or
460 maternal influence (Olin et al. 2011). A similar pattern was found in *R. taylori*,
461 although their FA niche size was comparatively small considering stomach content
462 analysis found a relatively broad range of species (e.g. prawns and teleost fishes)
463 (Simpfendorfer 1998). Interestingly, some *R. taylori* in Queensland occupied areas
464 near freshwater outflows during the wet season rather than moving to seagrass beds
465 with the majority of the population (Munroe et al. 2014). In light of these findings, it
466 may be that the narrow niche of *R. taylori* in the South Alligator estuary results from
467 this species targeting prey around freshwater outflows. Fatty acid niche metrics also
468 suggest that *R. taylori* consumes some riverine resources, however their prey
469 preference is largely associated with marine systems. Although FA niche metrics

470 indicate *R. taylori* and *G. garricki* share some resources, *G. garricki* mainly depend
471 on riverine resources with infrequent consumption of marine prey sources. Therefore
472 we suspect that *G. garricki* utilizes the lower South Alligator River whilst species
473 such as *R. taylori* and *C. amboinensis* take advantage of the diverse prey species in
474 the estuary.

475

476 In determining FA niche space, only five FAs could be used, which meant there was
477 an underlying assumption that these FAs represented all FAs within the species. A
478 more complete understanding of trophic niche based on FA profiles could be achieved
479 with larger sample sizes. Furthermore, a number of FAs were present in these species
480 either in relative high amounts or were the cause of separation between species that
481 are not yet associated with unique prey groups or food web source. A future area for
482 research is the identification of additional trophic markers to increase our ability to
483 distinguish food sources within and among rivers, estuaries and coastal ecosystems.

484

485 This study was the first to use SIA and FA analysis in combination to measure trophic
486 niche metrics and explore resource partitioning among an assemblage of consumers.
487 Our approach has demonstrated that elasmobranchs within the South Alligator River
488 display partitioning in trophic resource use, particularly across marine and freshwater
489 food webs that have sources related to seston, seagrass or macroalgae. Given the
490 potential for complexity in resource use, highlighted by seasonal shifts in isotopic
491 niches and the FA composition of muscle tissues, it appears likely these
492 elasmobranchs play important roles in food web connectivity in tropical aquatic
493 ecosystems. In particular, some species are utilizing food webs across the marine to
494 freshwater spectrum, which suggests that these fish provide broad and important

495 cross-biome trophic linkages in tropical coastal ecosystems. Differences in FA and SI
496 niche space highlighted the advantages of combining such analyses, which are likely
497 caused from the faster turnover of FAs, and more FA variables that may respond to
498 different trophic resource components. Further analysis to link these FA profiles with
499 potential prey items is needed to further increase our understanding of the role played
500 by elasmobranchs in coastal and estuarine ecosystems. Based on our findings, it is
501 concluded that tropical euryhaline and coastal elasmobranchs play important roles in
502 both middle- and higher-order trophic interactions across estuarine, coastal and
503 riverine ecosystems.

504

505

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520

521

522 **References**

523 Beckmann CL, Mitchell JG, Stone DAJ, Huveneers C (2014) Inter-tissue differences

524 in fatty acid incorporation as a result of dietary oil manipulation in Port Jackson

525 sharks (*Heterodontus portusjacksoni*). *Lipids* 49:577–590

526 Belicka LL, Matich P, Jaffé R, Heithaus MR (2012) Fatty acids and stable isotopes as

527 indicators of early-life feeding and potential maternal resource dependency in the

528 Bull Shark *Carcharhinus leucas*. *Mar Ecol Prog Ser* 455:245–256

529 Bligh EG, Dyer WJ (1959) A rapid method of total lipid extraction and purification.

530 *Can J Biochem Physiol* 37:911–917

531 Clarke K, Gorley R (2006) PRIMER v6 PRIMER-E Ltd. Plymouth, UK

532 Dalsgaard J, John MA St., Kattner G, Müller-Navarra D, Hagen W (2003) Fatty acid

533 trophic markers in the pelagic marine environment. *Adv Mar Biol* 46:225–340

534 Daly R, Froneman PW, Smale MJ (2013) Comparative feeding ecology of Bull

535 Sharks (*Carcharhinus leucas*) in the coastal waters of the southwest Indian

536 Ocean inferred from stable isotope analysis. *PLoS One* 8:1–11

537 Douglas MM, Bunn SE, Davies PM (2005) River and wetland food webs in

538 Australia's wet–dry tropics: general principles and implications for management.

539 *Mar Freshw Res* 56:329

540 Dulvy NK, Fowler SL, Musick JA, Cavanagh RD, Kyne PM, Harrison LR, Carlson

541 JK, Davidson LNK, Fordham SV, Francis MP, Pollock CM, Simpfendorfer CA,

542 Burgess GH, Carpenter KE, Compagno LJV, Ebert DA, Gibson C, Heupel MR,

543 Livingstone SR, Sanciangco JC, Stevens JD, Valenti S, White WT (2014)

- 544 Extinction risk and conservation of the world's sharks and rays. *Elife* 3:1–34
- 545 Every SL, Pethybridge HR, Crook DA, Kyne PM, Fulton CJ (In press) Comparison of
546 fin and muscle tissues for analysis of signature fatty acids in tropical euryhaline
547 sharks. *Journal Exp Mar Biol Ecol*
- 548 Froese R, Pauly D (2015) FishBase www.fishbase.org (accessed 20 Oct 2015)
- 549 Hall D, Lee SY, Meziane T (2006) Fatty acids as trophic tracers in an experimental
550 estuarine food chain: Tracer transfer. *J Exp Mar Bio Ecol* 336:42–53
- 551 Heithaus MR, Vaudo JJ, Kreicker S, Layman CA, Krützen M, Burkholder DA,
552 Gastrich K, Bessey C, Sarabia R, Cameron K, Wirsing AJ, Thomson JA,
553 Dunphy-Daly MM (2013) Apparent resource partitioning and trophic structure of
554 large-bodied marine predators in a relatively pristine seagrass ecosystem. *Mar*
555 *Ecol Prog Ser* 481:225–237
- 556 Hussey NE, Dudley SFJ, McCarthy ID, Cliff G, Fisk AT (2011) Stable isotope
557 profiles of large marine predators: viable indicators of trophic position, diet, and
558 movement in sharks? *Can J Fish Aquat Sci* 68:2029–2045
- 559 Hussey NE, Macneil MA, McMeans BC, Olin JA, Dudley SFJ, Cliff G, Wintner SP,
560 Fennessy ST, Fisk AT (2014) Rescaling the trophic structure of marine food
561 webs. *Ecol Lett* 17:239–250
- 562 Hussey NE, MacNeil MA, Olin JA, McMeans BC, Kinney MJ, Chapman DD, Fisk
563 AT (2012) Stable isotopes and elasmobranchs: tissue types, methods,
564 applications and assumptions. *J Fish Biol* 80:1449–1484
- 565 Hutchinson GE (1957) Concluding remarks. *Cold Spring Harb Symp Quant Biol*
566 22:415–427
- 567 Iverson SJ (2009) Tracing aquatic food webs using fatty acids: from qualitative
568 indicators to quantitative determination lipids in aquatic ecosystems. In: Kainz

- 569 M, Brett MT, Arts MT (eds) Lipids in Aquatic Ecosystems. Springer New York,
570 New York, NY, p 281–307
- 571 Jackson AL, Inger R, Parnell AC, Bearhop S (2011) Comparing isotopic niche widths
572 among and within communities: SIBER – Stable Isotope Bayesian Ellipses in R.
573 J Anim Ecol 80:595–602
- 574 Jayasinghe C, Gotoh N, Wada S (2003) Variation in lipid classes and fatty acid
575 composition of salmon shark (*Lamna ditropis*) liver with season and gender.
576 Comp Biochem Physiol Part B 134:287–295
- 577 Kelly J, Scheibling R (2012) Fatty acids as dietary tracers in benthic food webs. Mar
578 Ecol Prog Ser 446:1–22
- 579 Kim SL, Koch PL (2011) Methods to collect, preserve, and prepare elasmobranch
580 tissues for stable isotope analysis. Environ Biol Fishes 95:53–63
- 581 Kirsch P, Iverson S, Bowen W (1998) Dietary effects on the fatty acid signature of
582 whole Atlantic cod (*Gadus morhua*). Can J Fish Aquat Sci 55:1378–1386
- 583 Kitchell JF, Essington TE, Boggs CH, Schindler DE, Walters CJ (2002) The role of
584 sharks and longline fisheries in a pelagic ecosystem of the Central Pacific.
585 Ecosystems 5:202–216
- 586 Knip DM, Heupel MR, Simpfendorfer CA, Tobin AJ (2011) Ontogenetic shifts in
587 movement and habitat use of juvenile pigeye sharks *Carcharhinus amboinensis*
588 in a tropical nearshore region. Mar Ecol Prog Ser 425:233–246
- 589 Last PR, Stevens JD (2009) Sharks and Rays of Australia, Second Edition. CSIRO
590 Publishing, Collingwood
- 591 Layman CA, Araujo MS, Boucek R, Hammerschlag-Peyer CM, Harrison E, Jud ZR,
592 Matich P, Rosenblatt AE, Vaudo JJ, Yeager LA, Post DM, Bearhop S (2012)
593 Applying stable isotopes to examine food-web structure: an overview of

- 594 analytical tools. *Biol Rev Camb Philos Soc* 87:545–562
- 595 Layman CA, Arrington DA, Montaña CG, Post DM (2007) Can stable isotope ratios
596 provide for community-wide measures of trophic structure? *Ecology* 88:42–48
- 597 Logan JM, Lutcavage ME (2010) Stable isotope dynamics in elasmobranch fishes.
598 *Hydrobiologia* 644:231–244
- 599 Loneragan NR, Bunn SE, Kellaway DM (1997) Are mangrove and seagrasses sources
600 of organic carbon for penaeid prawns in a tropical estuary? A multiple isotope
601 study. *Mar Biol* 130:289–300
- 602 Lucifora LO, Carvalho MR de, Kyne PM, White WT (2015) Freshwater sharks and
603 rays. *Curr Biol* 25:R971–R973
- 604 Malpica-cruz L, Herzka SZ, Sosa-nishizaki O, Lazo JP (2012) Tissue-specific isotope
605 trophic discrimination factors and turnover rates in a marine elasmobranch:
606 empirical and modeling results. *Can J Fish Aquat Sci* 69:551–564
- 607 Martínez-Álvarez RM, Sanza A, García-Gallego M, Domezain A, Domezain J,
608 Carmonac R, M. del Valle, Ostos-Garrido AEM (2005) Adaptive branchial
609 mechanisms in the sturgeon *Acipenser naccarii* during acclimation to saltwater.
610 *Comp Biochem Physiol Biochem Physiol Part A* 141:183–190
- 611 Matich P, Heithaus MR (2014) Multi-tissue stable isotope analysis and acoustic
612 telemetry reveal seasonal variability in the trophic interactions of juvenile bull
613 sharks in a coastal estuary. *J Anim Ecol* 83:199–213
- 614 Matich P, Heithaus MR, Layman CA (2010) Size-based variation in intertissue
615 comparisons of stable carbon and nitrogen isotopic signatures of bull sharks
616 (*Carcharhinus leucas*) and tiger sharks (*Galeocerdo cuvier*). *Can J Fish Aquat*
617 *Sci* 67:877–885
- 618 Matich P, Heithaus MR, Layman CA (2011) Contrasting patterns of individual

- 619 specialization and trophic coupling in two marine apex predators. *J Anim Ecol*
620 80:294–305
- 621 McMeans BC, Arts MT, Lydersen C, Kovacs KM, Hop H, Falk-Petersen S, Fisk AT
622 (2013) The role of Greenland sharks (*Somniosus microcephalus*) in an Arctic
623 ecosystem: assessed via stable isotopes and fatty acids. *Mar Biol* 160:1223–1238
- 624 Munroe SEM, Heupel MR, Fisk AT, Logan JM, Simpfendorfer CA (2015) Regional
625 movement patterns of a small-bodied shark revealed by stable-isotope analysis. *J*
626 *Fish Biol* 86:1567–1586
- 627 Munroe SEM, Simpfendorfer CA, Heupel MR (2014) Habitat and space use of an
628 abundant nearshore shark, *Rhizoprionodon taylori*. *Mar Freshw Res* 65:959–968
- 629 Olin JA, Hussey NE, Fritts M, Heupel MR, Simpfendorfer CA, Poulakis GR, Fisk AT
630 (2011) Maternal meddling in neonatal sharks: implications for interpreting stable
631 isotopes in young animals. *Rapid Commun Mass Spectrom* 25:1008–16
- 632 Özogul Y, Özogul F, Alagoz S (2007) Fatty acid profiles and fat contents of
633 commercially important seawater and freshwater fish species of Turkey: A
634 comparative study. *Food Chem* 103:217–223
- 635 Papastamatiou YP, Wetherbee BM, Lowe CG, Crow LG (2006) Distribution and diet
636 of four species of carcharhinid shark in the Hawaiian Islands: evidence for
637 resource partitioning and competitive exclusion. *Mar Ecol Prog Ser* 320:239–251
- 638 Parrish CC (2013) Lipids in marine ecosystems. *ISRN Oceanogr* 2013:1–16
- 639 Parrish CC, Nichols PD, Pethybridge HR, Young JW (2015) Direct determination of
640 fatty acids in fish tissues: quantifying top predator trophic connections.
641 *Oecologia* 177:85–95
- 642 Parrish CC, Pethybridge HR, Young JW, Nichols PD (2013) Spatial variation in fatty
643 acid trophic markers in albacore tuna from the southwestern Pacific Ocean—A

- 644 potential “tropicalization” signal. Deep Sea Res Part II Top Stud Oceanogr:1–9
- 645 Peterson BJ, Fry B (1987) Stable isotopes in ecosystem studies. Annu Rev Ecol Syst
- 646 18:293–320
- 647 Pethybridge HR, Bodin N, Arsenault-Pernet E-J, Bourdeix J-H, Brisset B, Bigot J-L,
- 648 Roos D, Peter M (2014) Temporal and inter-specific variations in forage fish
- 649 feeding conditions in the NW Mediterranean: lipid content and fatty acid
- 650 compositional changes. Mar Ecol Prog Ser 512:39–54
- 651 Pethybridge HR, Daley RK, Nichols PD (2011) Diet of demersal sharks and
- 652 chimaeras inferred by fatty acid profiles and stomach content analysis. J Exp
- 653 Mar Bio Ecol 409:290–299
- 654 Pethybridge HR, Parrish CC, Morrongiello J, Young JW, Farley JH, Gunasekera RM,
- 655 Nichols PD (2015) Spatial patterns and temperature prediction of tuna fatty
- 656 acids: tracing essential nutrients and changes in primary producers. PLoS One
- 657 10:e0131598
- 658 Post DM, Layman CA, Arrington DA, Takimoto G, Quattrochi J, Montaña CG (2007)
- 659 Getting to the fat of the matter: models, methods and assumptions for dealing
- 660 with lipids in stable isotope analyses. Oecologia 152:179–189
- 661 Rooney N, McCann K, Gellner G, Moore JC (2006) Structural asymmetry and the
- 662 stability of diverse food webs. Nature 442:265–269
- 663 Ross ST (1986) Resource partitioning in fish assemblages: a review of field studies.
- 664 Copeia 2:352–388
- 665 Simpfendorfer CA (1998) Diet of the Australian Sharpnose Shark, *Rhizoprionodon*
- 666 *taylori*, from northern Queensland. Mar Freshw Res 49:757–761
- 667 Snelson FFJ, Mulligan TJ, Williams SE (1984) Food habits, occurrence, and
- 668 population structure of the Bull Shark, *Carcharhinus leucas*, in Florida coastal

- 669 lagoons. Bull Mar Sci 34:71–80
- 670 Speers-Roesch B, Ip YK, Ballantyne JS (2008) Plasma non-esterified fatty acids of
671 elasmobranchs: comparisons of temperate and tropical species and effects of
672 environmental salinity. Comp Biochem Physiol A Mol Integr Physiol 149:209-
673 216
- 674 Stevens JD, Bonfil R, Dulvy NK, Walker PA (2000) The effects of fishing on sharks,
675 rays, and chimaeras (chondrichthyans), and the implications for marine
676 ecosystems. ICES J Mar Sci 57:476–494
- 677 Swanson HK, Lysy M, Power M, Stasko AD, Johnson JD, Reist JD (2015) A new
678 probabilistic method for quantifying n-dimensional ecological niches and niche
679 overlap. Ecology 96:318–324
- 680 Syväranta J, Lensu A, Marjomäki TJ, Oksanen S, Jones RI (2013) An empirical
681 evaluation of the utility of convex hull and standard ellipse areas for assessing
682 population niche widths from stable isotope data. PLoS One 8:e56094
- 683 R Core Team(2014) R: A Language and Environment for Statistical Computing.
- 684 Thorburn DC, Gill HS, Morgan DL (2014) Predator and prey interactions of fishes of
685 a tropical Western Australia river revealed by dietary and stable isotope analyses.
686 J R Soc West Aust 97:363–387
- 687 Tillett BJ, Meekan MG, Field IC (2014) Dietary overlap and partitioning among three
688 sympatric carcharhinid sharks. Endanger Species Res 25:283–293
- 689 Tilley A, López-Angarita J, Turner JR (2013) Diet reconstruction and resource
690 partitioning of a Caribbean marine mesopredator using stable isotope Bayesian
691 modelling. PLoS One 8:e79560
- 692 Uysal K, Aksoylar MY (2005) Seasonal variations in fatty acid composition and the
693 n-6/n-3 fatty acid ratio of pikeperch (*Sander lucioperca*) muscle lipids. Ecol

694 Food Nutr 44:23–35

695 Warfe DM, Pettit NE, Davies PM, Pusey BJ, Hamilton SK, Kennard MJ, Townsend
696 SA, Bayliss P, Ward DP, Douglas MM, Burford MA, Finn M, Bunn SE,
697 Halliday IA (2011) The “wet-dry” in the wet-dry tropics drives river ecosystem
698 structure and processes in northern Australia. *Freshw Biol* 56:2169–2195

699 White WT, Appleyard SA, Sabub B, Kyne PM, Harris M, Lis R, Baje L, Usu T,
700 Smart JJ, Corrigan S, Yang L, Naylor GJP (2015) Rediscovery of the threatened
701 river sharks, *Glyphis garricki* and *G. glyphis*, in Papua New Guinea. *PLoS One*
702 10:e0140075

703 Winemiller KO, Jepsen DB (1998) Effects of seasonality and fish movement on
704 tropical river food webs. *J Fish Biol* 53:267–296

705 Yick JL, Tracey SR, White RWG (2011) Niche overlap and trophic resource
706 partitioning of two sympatric batoids co-inhabiting an estuarine system in
707 southeast Australia. *Appl Ichthyology* 27:1272–1277

708

709 **Tables and Figures**710 **Table 1.** Number (*n*) and total length (TL) of seven euryhaline and coastal

711 elasmobranch species from the South Alligator River, Kakadu National Park,

712 Australia from which muscle tissue samples were taken for stable isotope (SIA) and

713 fatty acid analysis (FAA).

Species		<i>n</i> (Wet:Dry)		Range TL (cm)		Sex ratio (M:F)	
Scientific Name	Common Name	SIA	FAA	SIA	FAA	SIA	FAA
<i>Carcharhinus amboinensis</i>	Pigeye Shark	0:2	0:1	68-68	68	1:1	0:1
<i>Carcharhinus leucas</i>	Bull Shark	28:6	2:20	93-72	72-139	18:2	9:1
<i>Glyphis garricki</i>	Northern River Shark	22:20	12:13	56-141	45-61	22:2	15:1
<i>Glyphis glyphis</i>	Speartooth Shark	2:3	2:6	71-120	78-136	1:4	6:2
<i>Himantura dalyensis</i>	Freshwater Whipray	0:2	0:2	56 -110	56-110	0:2	0:2
<i>Pristis pristis</i>	Largetooth Sawfish	2:0	0	87-96	-	1:1	
<i>Rhizoprionodon taylori</i>	Australian Sharpnose Shark	8:30	28:1	34-87	34 -87	11:3	6:2

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724 **Table 2.** Mean values for fatty acids (FA > 0.5%), $\delta^{13}\text{C}$ and $\delta^{14}\text{N}$ and their standard

725 deviation in muscle tissue from seven elasmobranchs collected from the South

726 Alligator River, Kakadu National Park, Australia. Included are calculations for

727 trophic position (TP), total area of stable isotope convex hulls (TA), ellipse area of SI

728 (SEAc) and major EFA niche space.

SI	<i>C. amboinensis</i>	<i>C. leucas</i>	<i>G. garricki</i>	<i>G. glyphis</i>	<i>H. dalyensis</i>	<i>P. pristis</i>	<i>R. taylori</i>
$\delta^{13}\text{C}$	-13.4	-14.9±2.3	-18.9±1.3	-18.9±1.7	-20.2	-13.4	-15.1±1.3
$\delta^{15}\text{N}$	11.7	12.95±2.8	9.3±1.9	7.6±3.6	8.9	12.6	11.1±2.5
C:N	2.6	2.8±0.2	2.8±0.2	2.7±0.1	3.2	2.7	2.8±0.4
TP	4.2	4.8±0.9	3.6±0.4	3.2±0.8	3.4	4.5	4.1±0.6
SEAc (Wet/Dry)		18.5 (16.9/17.1)	6.9 (9.8/7.0)				10.3 (2.1/11.7)
TA (Wet/Dry)		79.1 (17.8/65.8)	28.9 (21.3/37.8)				39.1 (2.7/39.1)
FA							
18:1 ω 9	14.1	16.2±6.3	11.2±4.9	8.2±3.6	13.8±0.5		9.5±2.7
18:2b ν	0.6	0.6±0.4	0.2±0.2	0.5±0.2	0.4±0.3		0.1±0.1
18:2c ν	0.1	0.9±0.8	0.1±0.3	0.1±0.1	0.1±0.1		0.13±0.02
18:2 ω 6	0.6	0.8±0.9	1.8±1.1	1.8±1.1	3.8±3.7		0.9±1.0
20:2 ω	0.2	2.8±2.2	0.5±0.8	0.3±0.2	0.2±0.1		0.1±0.1
20:2 ω 6	0.4	0.5±0.8	0.9±0.4	0.8±0.2	0.3±0.0		0.6±0.2
20:3 ω 9 $\#$	0.8	7.6±6.2	1.4±3.4	0.3±0.3	0.3±0.1		0.1±0.1
20:3 ω 6	0.2	0.4±0.3	0.8±0.4	0.6±0.3	0.3±0.1		0.4±0.2
22:3 ω	1.7	1.01±0.7	1.7±1.8	2.6±2.4	2.9±3		1.1±1.2
20:4 ω 6	9.6	5.3±5.5	11.3±4.7	14.1±4.4	14.4±2.4		8.1±2.1
22:4 ω 6	3.5	2.1±2.1	6.4±4.2	8.0±6.5	0.2±0.2		3.0±1.4
20:5 ω 3	0.9	0.5±0.2	0.9±0.8	1.1±0.6	5.2±6.8		1.6±0.8
22:5 ω 3	0.9	1.9±1.2	1.5±2.0	1.6±1.8	1.1±1.6		1.1±1.7
22:5 ω 6	1.8	1.2±0.7	2.3±1.1	2.3±1.1	1.2±0.8		2.2±0.8
22:6 ω 3	4.8	4.8±2.3	8.3±4.6	9.8±8.8	4.3±0.2		11.4±5.6
ω 3/ ω 6	0.4	0.70±0.4	0.5±0.6	0.4±0.8	0.5±0.0		0.9±1.4
EFA niche		13413.7	114886.8				455.6
EFA Niche Dry/Wet			40732.8/ 64824.2				

729 EFAs (Essential fatty acids) <0.5% include: 18:2a ν , 18:4 ω 3, 18:3 ω 6, 18:3 ω 3, 20:4 ω 3/20:2, 21:5 ω 3,

730 21:3, 22:2a ν , and 22:2b ν .

731 # 20:3 ω 9 identified based on comparison with other *C. leucas* fatty acid literature; a standard was not

732 available at the time of analyses. ν = unable to identify bonds as standard was not available at the time

733 of analyses.

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Species	Variable	Source	DF	F-ratio	P -Value	Residual standard Error	R ²	t-value	Significant Tukey post hoc comparisons (p<0.05)
All	Species	δ ¹⁵ N	115	16.59	< 0.01	2.5	0.3		<i>G. garricki</i> – <i>C. leucas</i> <i>G. glyphis</i> – <i>C. leucas</i> <i>R. taylori</i> – <i>G. garricki</i> <i>R. taylori</i> – <i>G. glyphis</i> <i>G. garricki</i> – <i>C. leucas</i> <i>G. glyphis</i> – <i>C. leucas</i> <i>R. taylori</i> – <i>G. garricki</i> <i>R. taylori</i> – <i>C. leucas</i> <i>R. taylori</i> – <i>G. glyphis</i> <i>G. garricki</i> – <i>C. leucas</i> <i>G. glyphis</i> – <i>C. leucas</i>
		δ ¹³ C	115	50.7	< 0.01	1.7	0.6		
		ω3/ω6	3 & 76	7.2	< 0.01	0.6	0.22		
<i>C. leucas</i>	Season	δ ¹⁵ N	30	0.8	0.4	2.6	0.02		
		δ ¹³ C	30	5.4	0.03	1.9	0.15		
		Season + TL	29	5.1	0.01	1.8	0.26		
<i>G. garricki</i>	Season	δ ¹⁵ N	40	0.02	0.9	1.9	0.0		
		δ ¹³ C	40	1.3	0.2	1.3	0.03		
<i>R. taylori</i>	Season	δ ¹⁵ N	36	1.4	0.2	2.5	0.04		
		δ ¹³ C	36	1.7	0.2	1.3	0.04		
	Season + TL	δ ¹³ C			0.7			0.4	
	Season + TL	δ ¹³ C			0.05			-2.05	

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736 **Table 3.** ANOVA and ANCOVA (shaded) of stable isotope values and the ratio of

737 ω3/ω6 from the muscle of *Carcharhinus leucas*, *Glyphis garricki*, *G. glyphis* and

738 *Rhizoprionodon taylori* from the South Alligator River, Kakadu National Park,

739 Australia compared to the “wet” and “dry” season, with total length (TL) the

740 covariate. Significant values in bold.

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750 **Table 4.** Isotopic overlap (%) of stable isotopes (SI) ellipses (SEAC) and convex hull
 751 of *Carcharhinus leucas*, *Glyphis garricki* and *Rhizoprionodon taylori* from the South
 752 Alligator River, Kakadu National Park, Australia. Also included are the probabilities
 753 of sharks being in the major essential fatty acids (EFA) niche space of each other with
 754 a confidence interval (CI) of 95%.

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Species comparisons	% Overlap ellipse (SI)			% Overlap convex hull (SI)			EFA Ellipses
	All Seasons	Dry Season	Wet Season	All Seasons	Dry Season	Wet Season	% Probability (95% CI)*
<i>C. leucas</i> x <i>G. garricki</i>							61.7
<i>C. leucas</i> x <i>R. taylori</i>							0.4
<i>G. garricki</i> x <i>C. leucas</i>	0	19.4	0	20.8	18.3	16.8	13.2
<i>G. garricki</i> x <i>R. taylori</i>	0	0	0	12.4	6.5	0	0.6
<i>R. taylori</i> x <i>C. leucas</i>	23.7	16.4	10.9	31.3	20.6	4.0	11.2
<i>R. taylori</i> x <i>G. garricki</i>							79.9
Season comparisons							
<i>G. garricki</i> (Dry) x (Wet)		9.5			33.8		50.3
<i>G. garricki</i> (Wet) x (Dry)							35.4
<i>C. leucas</i> (Dry) x (Wet)		18.3			20.1		
<i>R. taylori</i> (Dry) x (Wet)		12.0			6.5		

756 * of 1 species being in another's niche space.

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759 **Figure captions**

760 **Fig. 1** Map of the South Alligator River, Northern Territory, Australia, showing
761 capture locations of elasmobranchs. Inset shows where the river is in relation to
762 northern Australia. (Background map from Stamen Toner, and formed using QGIS
763 Development Team, (2015)).

764 **Fig. 2** Biplot of mean $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ (\pm standard deviations) values for seven species
765 of euryhaline and coastal elasmobranchs from South Alligator River, Kakadu
766 National Park, Australia Primary producer values from the wet and dry season,
767 Embley River Estuary, Gulf of Carpentaria (Loneragan et al., 1997).

768 **Fig. 3** (a) Bivariate plot of isotopic space depicting wet vs dry niche areas within
769 standard ellipses of $\delta^{13}\text{C}$ and $\delta^{15}\text{N}$ of *Carcharhinus leucas* (blue) *Glyphis garricki*
770 (black), and *Rhizoprionodon taylori* (red) from Kakadu National Park, Australia (b)
771 Bayesian confidence intervals of isotopic niche area.

772 **Fig. 4** Principal component analysis of fatty acids in muscle tissue of *Carcharhinus*
773 *leucas*, *Glyphis garricki* and *Rhizoprionodon taylori*, *Carcharhinus amboinensis* and
774 *Himantura dalyensis* from the South Alligator River, Kakadu National Park,
775 Australia.

776 **Fig. 5** Comparisons of the posterior distributions of the probabilistic niche overlap
777 metrics (%) for a specified niche region of 95 % that an individual shark from the
778 species in the row will be found in the niche of *Carcharhinus leucas* (blue), *Glyphis*
779 *garricki* (black) (wet (black) and dry season (grey)) and *Rhizoprionodon taylori* (red)
780 of 5 major essential fatty acids (EFA). The posterior means and 95% credible
781 intervals are displayed in light blue.

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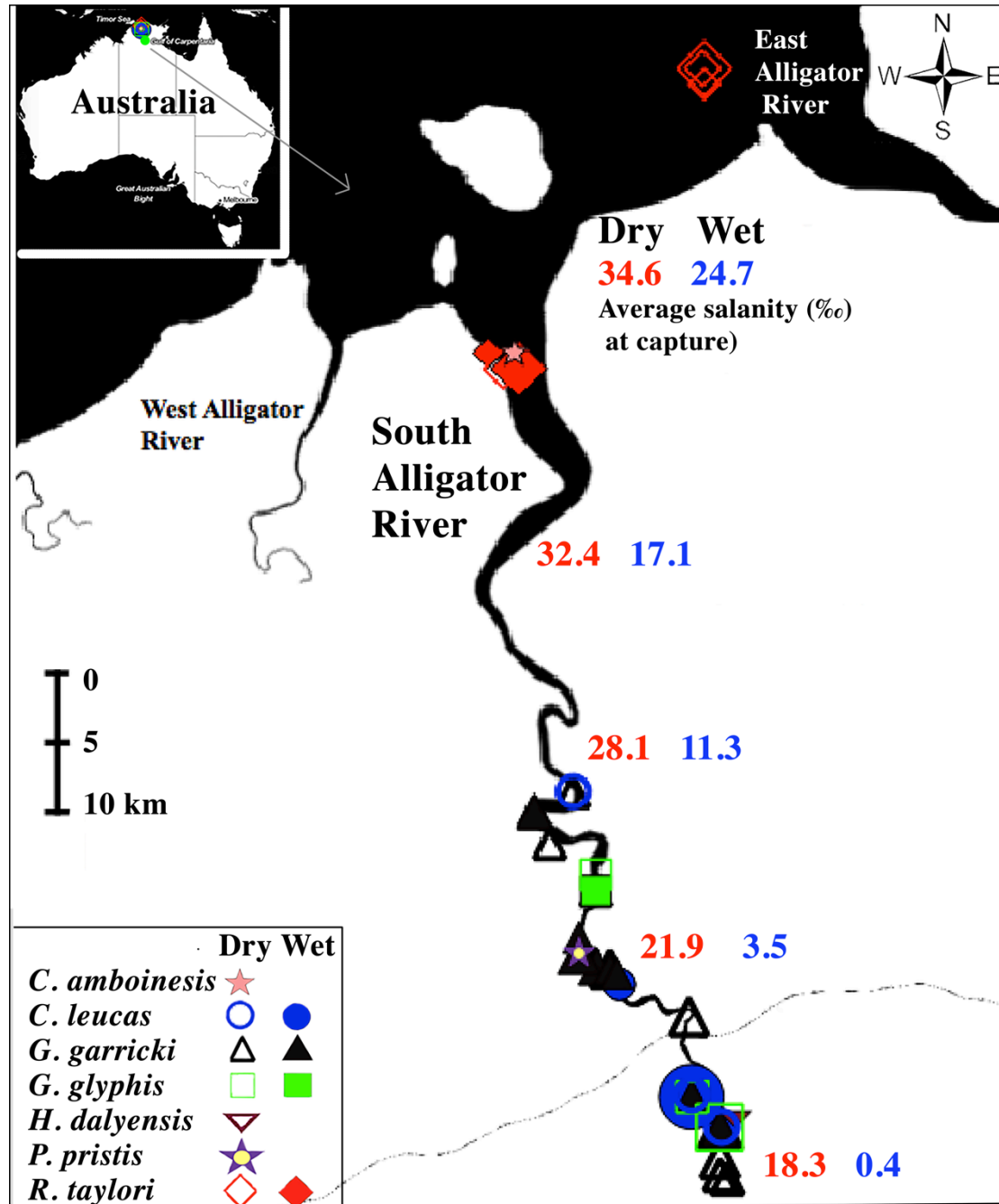
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Fig. 1



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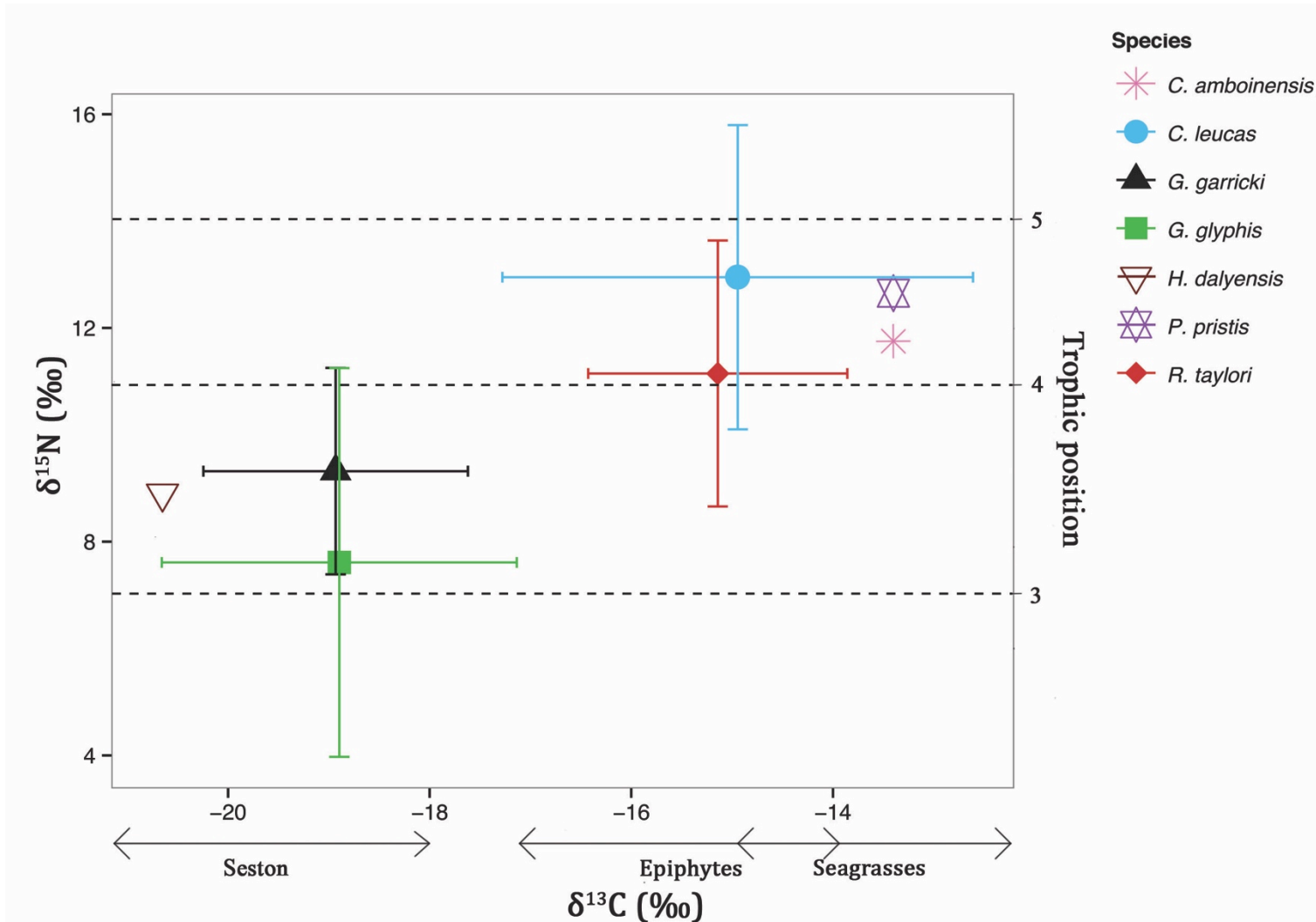
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792 **Fig. 2**

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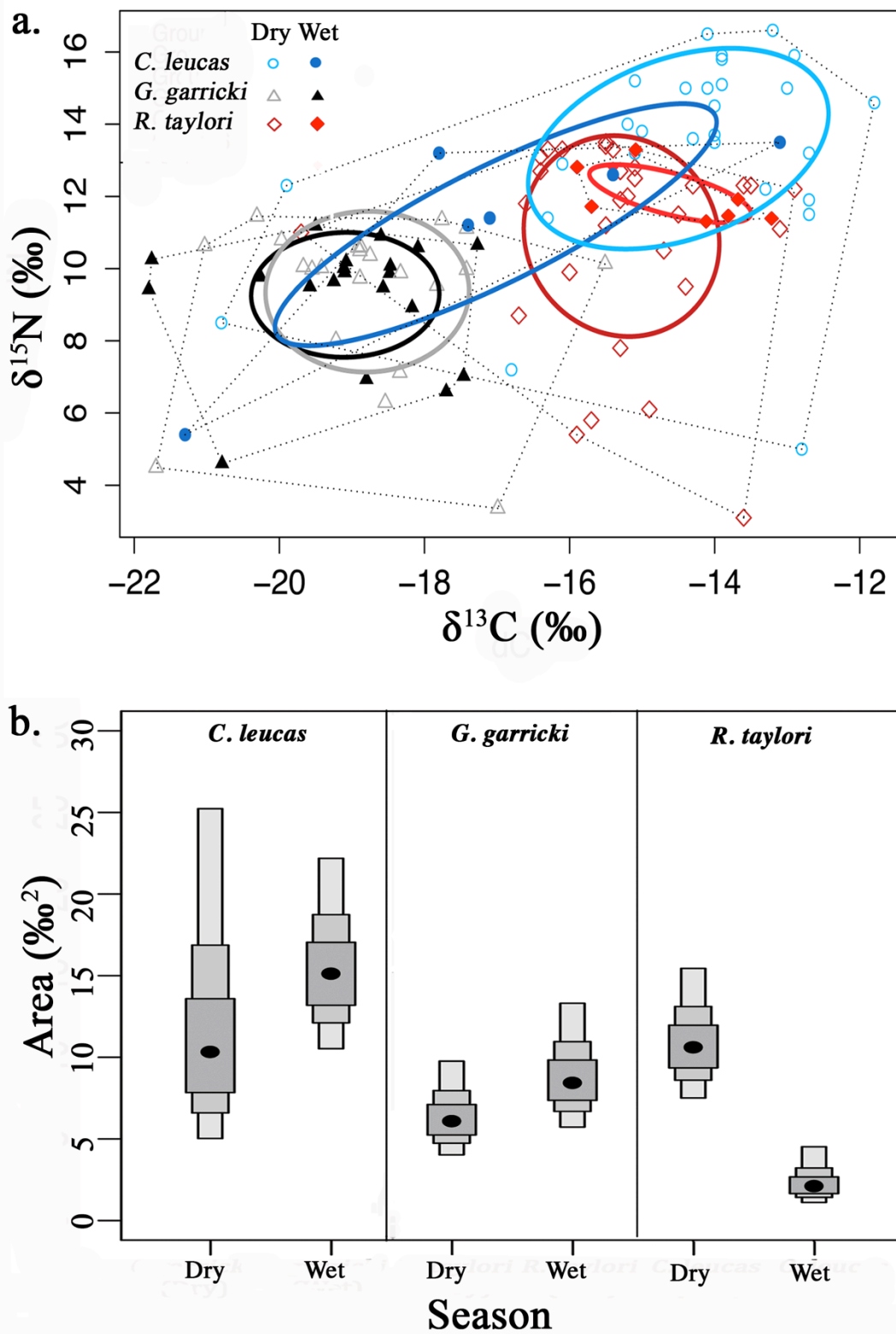
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805 **Fig. 3**



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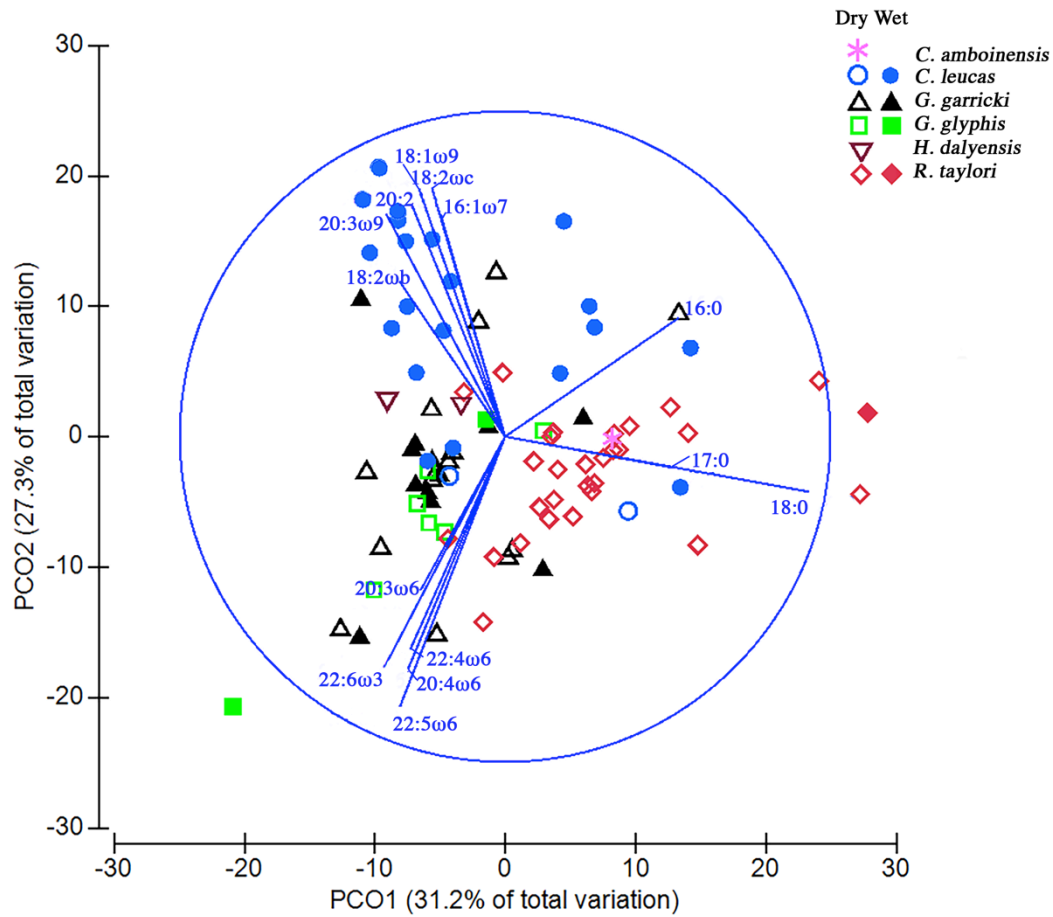
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808 **Fig. 4**

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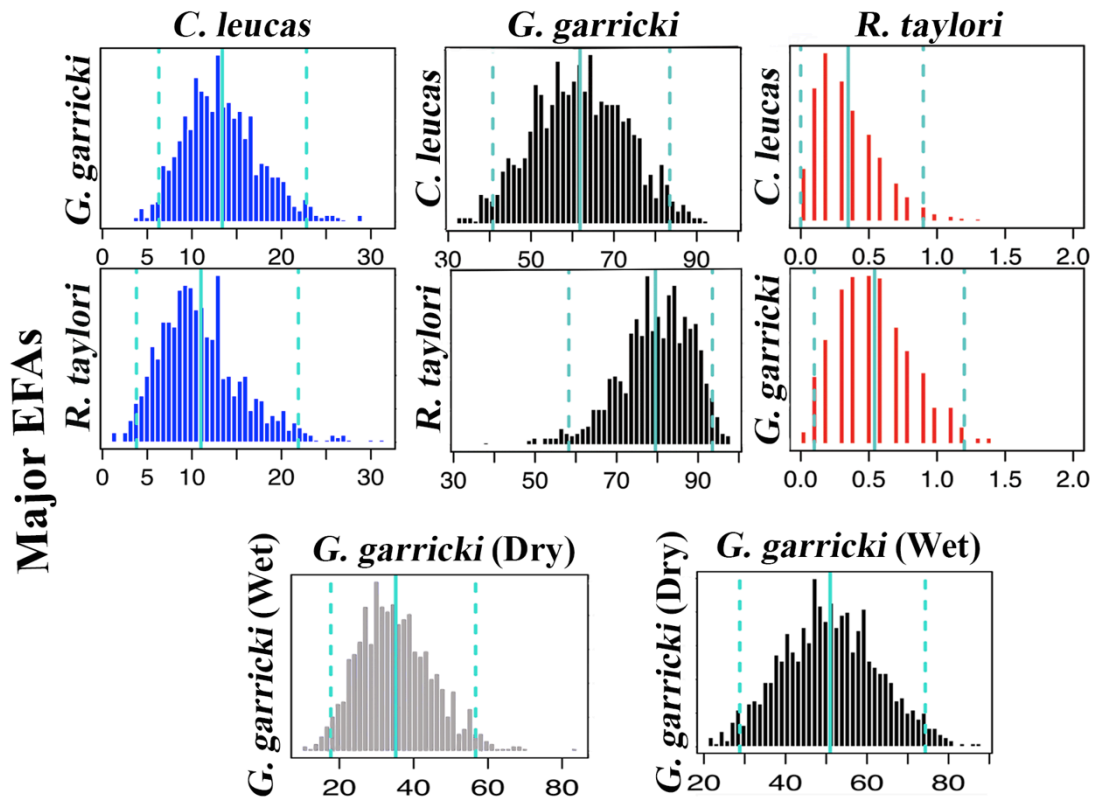
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813 **Fig. 5**

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Overlap Probability -- Niche Region Size: 95%

815 **Supplementary 1 Table S1**

816 Mean values for all fatty acids (FA > 0.5%), and their standard deviation in muscle

817 tissue from six elasmobranchs collected from the South Alligator River, Kakadu

818 National Park, Australia.

FAs	<i>C. amboinensis</i>	<i>C. leucas</i>	<i>G. garricki</i>	<i>G. glyphis</i>	<i>H. dalyensis</i>	<i>R. taylori</i>
16:0	8.6	10.6±4.8	8.0±5.6	8.0±5.5	13.7±3.5	13.5±5.1
17:0	0.6	0.4±0.3	0.6±0.2	0.68±0.4	1.1±0.0	0.9±0.2
18:0	32.5	19.8±6.9	19.5±5.2	17.1±4.5	13.8±5.8	28.3±7.4
22:0	0.3	0.6±0.4	1.7±2.9	0.8±0.6	0.8±0.4	0.5±0.6
24:0	0.5	0.4±0.3	0.5±0.4	1.0±0.9	0.7±0.7	0.4±0.3
15:1	0.2	1.3±1.3	1.5±1.5	0.8±0.8	1.1±0.3	0.7±0.4
16:1ω7	0.5	1.9±1.36	0.7±0.6	0.7±0.5	1.6±0.2	0.7±0.5
17:1	1.3	1.3±0.9	1.0±0.3	2.0 ±1.7	3.8±0.7	0.5±0.2
18:1ω9	14.1	16.2±6.3	11.2±4.9	8.2±3.6	13.8±0.5	9.5±2.7
18:1ω7	6.2	4.9±2.5	5.2±1.8	4.3±2.0	4.0±1.3	7.2±2.0
17:1ω6	0.5	0.5±0.2	0.5±0.3	0.6±0.5	0.8±0.4	0.6±0.3
20:1ω9	0.8	1.0±0.4	1.0±0.5	0.8±0.7	0.3±0.3	0.7±0.3
22:1ω11	0.1	1.7±5.0	0.2±0.2	0.1±0.0	0.1±0.1	0.5±2.1
24:1ω9	0.9	0.8±0.4	0.9±0.7	1.2±0.9	0.6±0.9	0.4±0.1
18:2b^v	0.6	0.6±0.4	0.2±0.2	0.5±0.2	0.4±0.3	0.1±0.1
18:2c^v	0.1	0.9±0.8	0.1±0.3	0.1±0.1	0.1±0.1	0.13±0.02
18:2ω6	0.6	0.8±0.9	1.8±1.1	1.8±1.1	3.8±3.7	0.9±1.0
20:20	0.2	2.8±2.2	0.5±0.8	0.3±0.2	0.2±0.1	0.1±0.1
20:2ω6	0.4	0.5±0.8	0.9±0.4	0.8±0.2	0.3±0.0	0.6±0.2
20:3ω9[#]	0.8	7.6±6.2	1.4±3.4	0.3±0.3	0.3±0.1	0.1±0.1
20:3ω6	0.2	0.4±0.3	0.8±0.4	0.6±0.3	0.3±0.1	0.4±0.2
22:3	1.7	1.01±0.7	1.7±1.8	2.6±2.4	2.9±3	1.1±1.2
20:4ω6	9.6	5.3±5.5	11.3±4.7	14.1±4.4	14.4±2.4	8.1±2.1
22:4ω6	3.5	2.1±2.1	6.4±4.2	8.0±6.5	0.2±0.2	3.0±1.4
20:5ω3	0.9	0.5±0.2	0.9±0.8	1.1±0.6	5.2±6.8	1.6±0.8
22:5ω3	0.9	1.9±1.2	1.5±2.0	1.6±1.8	1.1±1.6	1.1±1.7
22:5ω6	1.8	1.2±0.7	2.3±1.1	2.3±1.1	1.2±0.8	2.2±0.8
22:6ω3	4.8	4.8±2.3	8.3±4.6	9.8±8.8	4.3±0.2	11.4±5.6
Σ<5% FAs	0.1	5.5±0.9	5.6±0.9	6.1±1.0	5.1±0.8	3.03±0.5
i17:0	0.6	0.8±0.5	1.0±2.8	0.6±0.4	0.7±0.6	0.3±0.1
16:0FALD	0.6	0.8±0.7	1.1±1.0	1.1±1.3	0.9±0.6	0.76±0.6
18:0FALD	1.6	0.9±0.6	1.3±1.1	1.7±1.3	2.2±1.4	0.5±0.2

819 FAs <0.5% include 14:0 15:0, a15:0, 15:0, 14:1, 16:1ω13, 16:1ω9, 16:1ω7, 16:1ω5, 17:1ω8+a17:0,

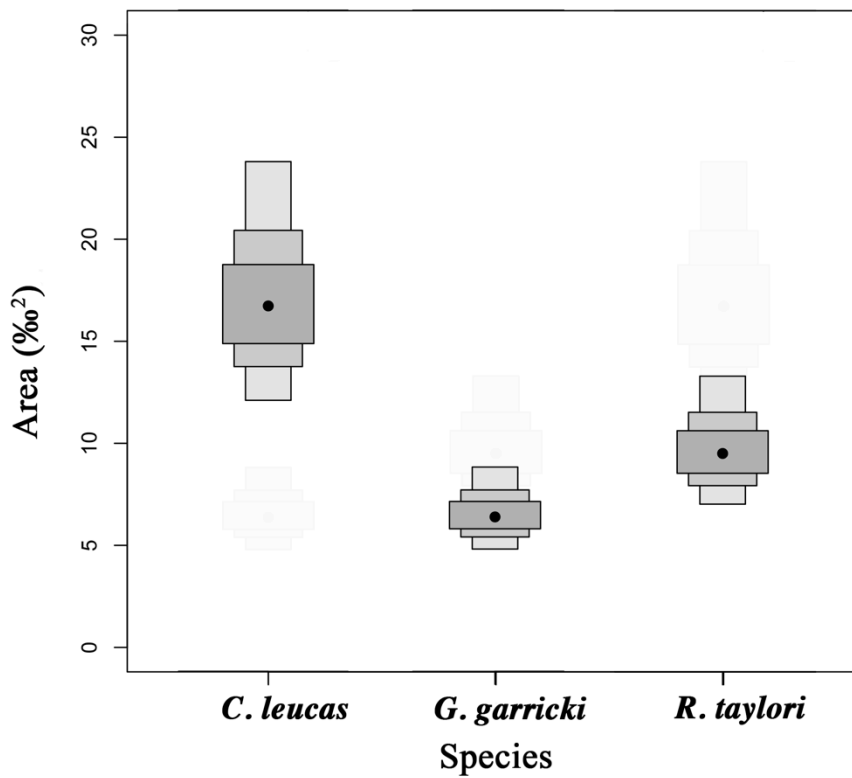
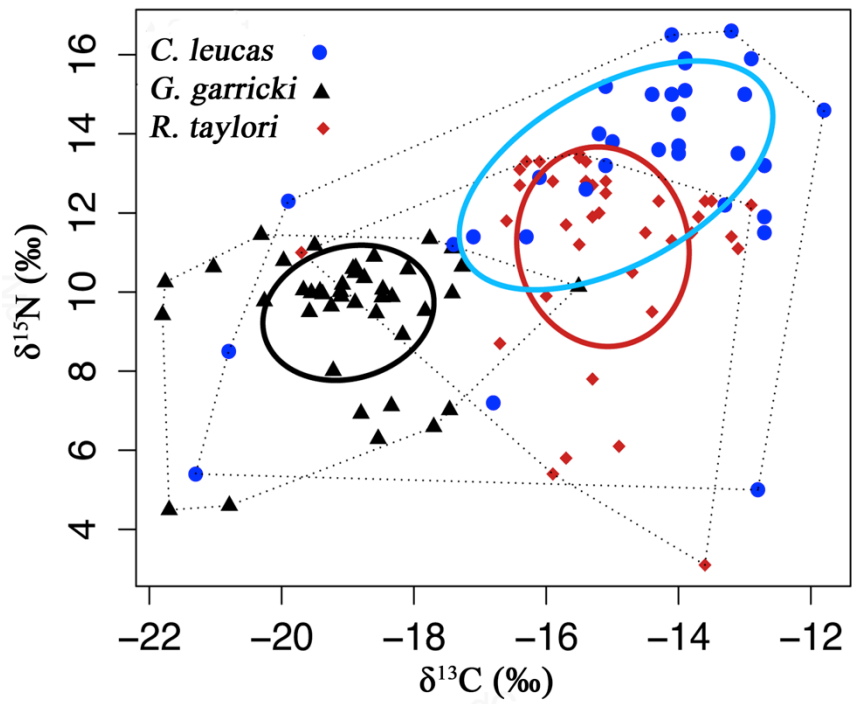
820 18:1ω7, 18:1ω5, 18:1, 19:1, 20:1ω7, 20:1ω11, 20:1ω5, 22:1ω9, 22:1ω7, 24:1ω11, 24:1ω7, 16:4+16:3,

821 18:2a^v, 18:4ω3, 18:3ω6, 18:3ω3, 20:4ω3/20:2, 21:5ω3, 21:3, 22:2a^v, 22:2b^v, i16:0, 18:1FALD

822 # 20:3 ω 9 identified based on comparison with other *C. leucas* fatty acid literature; a standard was not
823 available at the time of analyses. ^v= unable to identify bonds as standard was not available at the time
824 of analyses. FA - Fatty acids, SAT- saturated fatty acids, MUFA - monounsaturated fatty acids, PUFA
825 - polyunsaturated fatty acids. FALD – fatty aldehyde analyzed as dimethyl acetal.
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Supplementary 2 – Fig. S1

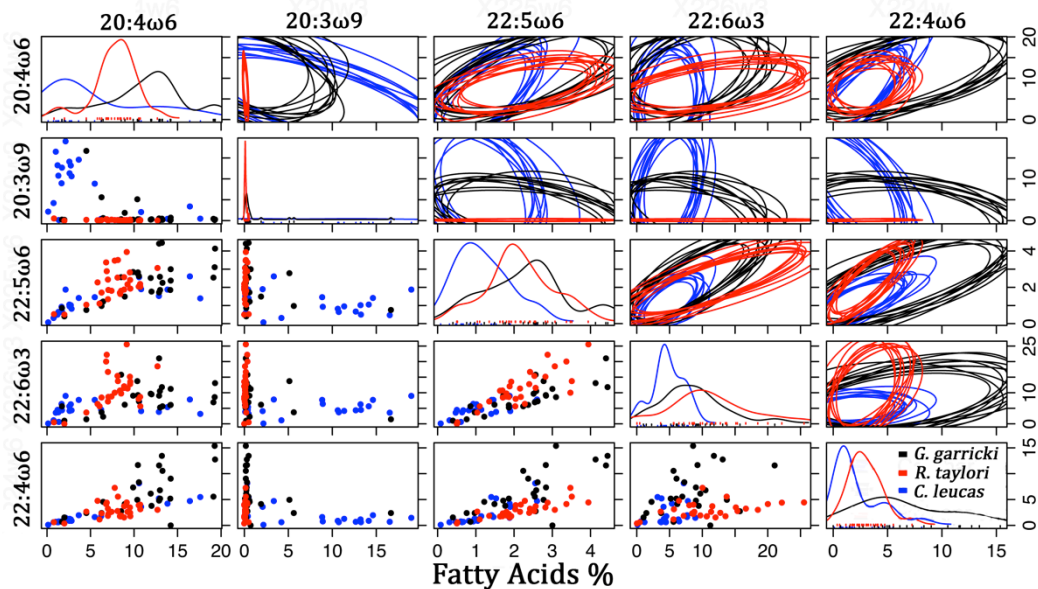


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831 **Supplementary 3 – Fig. S2**

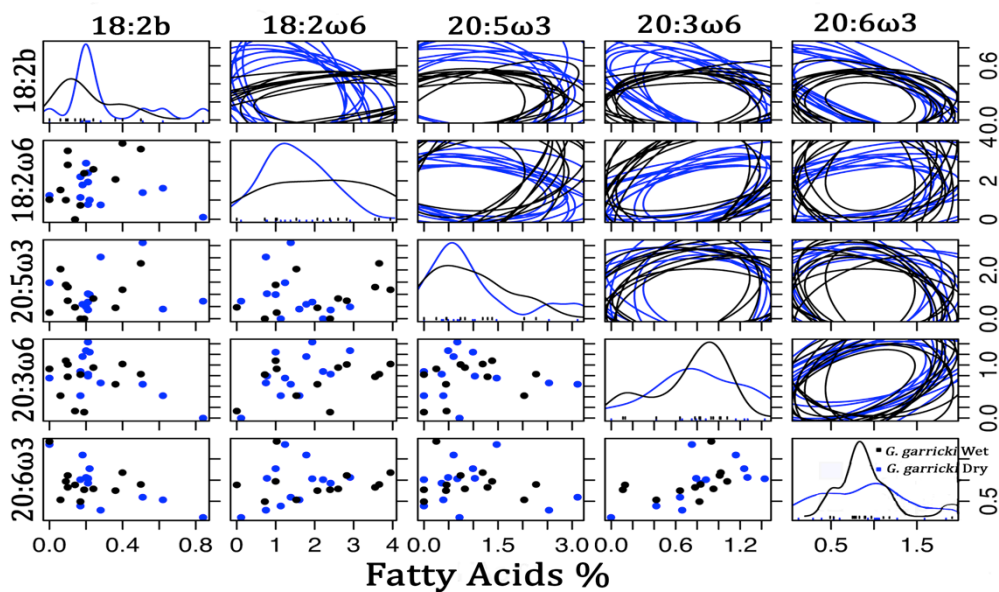
832 (i) Ten random elliptical projections of trophic niche region for each elasmobranch
833 and pair of major essential fatty acids (FA) and (ii) *G. garricki* wet vs dry season
834 major (elliptical plots). Also displayed are one-dimensional density plots (lines) and
835 two-dimensional scatterplots to demonstrate normality.

836 **a.**



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b.



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